

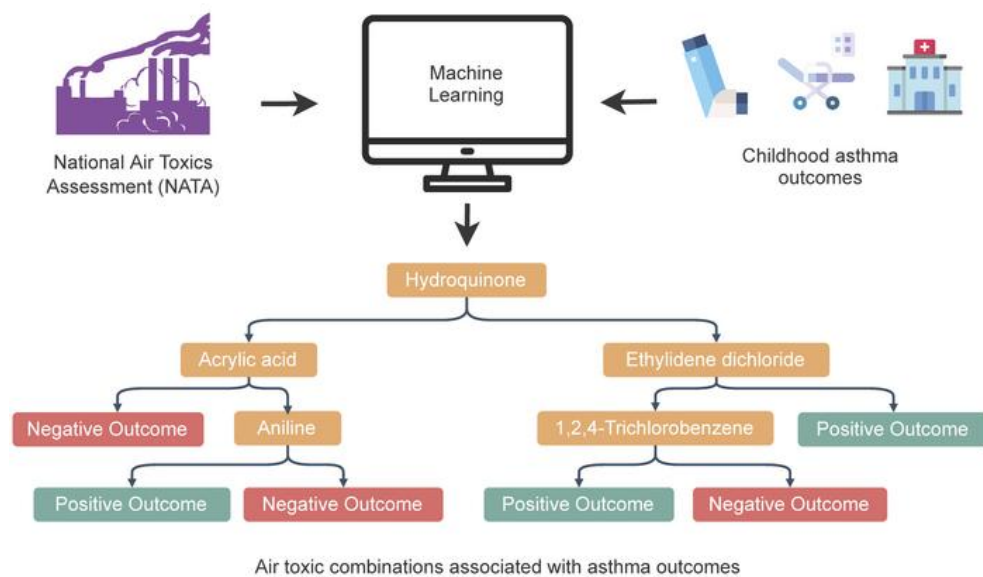
## Machine learning-driven identification of early-life air toxic combinations associated with childhood asthma outcomes

Yan-Chak Li, ... , Gaurav Pandey, Supinda Bunyavanich

*J Clin Invest.* 2021. <https://doi.org/10.1172/JCI152088>.

Research In-Press Preview Pulmonology

### Graphical abstract



Find the latest version:

<https://jci.me/152088/pdf>



1 **Title:** Machine learning-driven identification of early-life air toxic combinations associated  
2 with childhood asthma outcomes

3

4 **Authors:** Yan-Chak Li<sup>1§</sup>, Hsiao-Hsien Leon Hsu<sup>2,3§</sup>, Yoojin Chun<sup>1</sup>, Po-hsiang Chiu<sup>1</sup>, Zoe Arditì<sup>1,4</sup>,  
5 Luz Claudio<sup>2,3</sup>, Gaurav Pandey<sup>1,3\*</sup>, Supinda Bunyavanich<sup>1,4\*</sup>

6

7 **Affiliations:**

8 <sup>1</sup>Department of Genetics and Genomic Sciences, Icahn School of Medicine at Mount Sinai  
9 and Icahn Institute for Data Science and Genomic Technology, New York, NY, USA

10 <sup>2</sup>Department of Environmental Medicine and Public Health, Icahn School of Medicine at  
11 Mount Sinai, New York, NY, USA

12 <sup>3</sup>Institute for Exposomic Research, Icahn School of Medicine at Mount Sinai, New York, NY,  
13 USA

14 <sup>4</sup>Division of Allergy and Immunology, Department of Pediatrics, Icahn School of Medicine at  
15 Mount Sinai, New York, NY, USA

16 <sup>§</sup>Contributed equally to this work

17 <sup>\*</sup>Co-corresponding Authors

18

- 19 **Correspondence:** Supinda Bunyavanich, One Gustave Levy Place, Box #1498, New York, NY  
20 10029, tel. +1-212-241-5548, [supinda@post.harvard.edu](mailto:supinda@post.harvard.edu); Gaurav Pandey, One Gustave Levy  
21 Place, Box #1498, New York, NY 10029, tel. +1-212-659-8535, [gaurav.pandey@mssm.edu](mailto:gaurav.pandey@mssm.edu)

22 **Abstract:**

23 Air pollution is a well-known contributor to asthma. Air toxics are hazardous air pollutants  
24 that cause or may cause serious health effects. While individual air toxics have been  
25 associated with asthma, only a limited number of studies have specifically examined  
26 combinations of air toxics associated with the disease. We geocoded air toxic levels from  
27 the US National Air Toxics Assessment (NATA) to residential locations for participants of our  
28 AiRway in Asthma (ARIA) study. We then applied **Data-driven Exposure Profile** extraction  
29 (DEEP), a novel machine learning-based method, to discover combinations of early-life air  
30 toxics associated with current use of daily asthma controller medication, lifetime emergency  
31 department visit for asthma, and lifetime overnight hospitalization for asthma. We  
32 discovered 20 multi-air toxic combinations and 18 single air toxics associated with at least  
33 one outcome. The multi-air toxic combinations included those containing acrylic acid,  
34 ethylidene dichloride, and hydroquinone, and they were significantly associated with  
35 asthma outcomes with odds ratios of 1.60 to 3.19. Several air toxic members of the  
36 combinations would not have been identified by single air toxic analyses, supporting the use  
37 of machine learning-based methods designed to detect combinatorial effects. Our findings  
38 provide knowledge about air toxic combinations associated with childhood asthma.

## 39 **Introduction**

40 Air toxics are hazardous air pollutants that cause or may cause serious health effects

41 (1). They are well-established detriments to human respiratory health, especially for

42 children (2-8). In particular, exposure to air toxics early in life predisposes children to

43 asthma, one of the most prevalent diseases in this demographic group. Epidemiologic

44 studies have linked prenatal and early life exposure to air toxics with childhood wheeze,

45 asthma, and altered lung function (6-14).

46 Although air toxics are generally analyzed and regulated as individual chemicals (6), we

47 are exposed to combinations of air toxics in ambient air. The specific combinations of

48 individual air toxics that influence childhood asthma have not been studied adequately.

49 Assessing the respiratory health effects of multiple air toxics is challenging for several

50 reasons (7, 15). First, it is logistically difficult and expensive to collect detailed individualized

51 exposure data for multiple air toxics using personal or local monitoring. Additionally, there

52 are limited statistical methods to parse the effects of mixtures where individual air toxics

53 may only contribute slightly to an adverse outcome, but have a different impact in

54 combination with other air toxics (15). As a result, few studies have considered

55 simultaneous exposure to air toxic mixtures and their associations with children's health,

56 including asthma (8, 15-17).

57           Several studies linking air toxic mixtures and health outcomes, as well as a prior review  
58 of 57 studies that examined air pollutants and their health effects, reached no consensus on  
59 the ideal methods for multi-pollutant analyses (6, 7, 15, 16, 18). A key limitation of the  
60 studies reviewed was that most metrics assumed pure additivity of the effects of multiple  
61 air toxics, without consideration of synergistic and/or antagonistic interactions. Due to these  
62 challenges, air toxic combinations that collectively influence childhood asthma remain  
63 suboptimally characterized. Furthermore, identifying air toxic combinations associated with  
64 health outcomes is also difficult due to the exponentially large number of combination  
65 subsets in a set of air toxics, i.e.,  $2^N-1$  combinations in a set of N air toxics. Conventional  
66 statistical methods (19-24) and feature importance assessment using machine learning  
67 algorithms (16, 25-27) have not been effective for this task since they generally assess the  
68 association of air toxics individually.

69           In this study, we hypothesized that exposure to combinations of air toxics during early  
70 life is associated with asthma outcomes in later childhood. These outcomes included current  
71 need for daily asthma controller medication, lifetime emergency room visit for asthma, and  
72 lifetime overnight hospitalization for asthma (**Figure 1**). Asthma-related medication use,  
73 emergency room visits, and hospitalizations are frequently studied asthma outcomes that  
74 reflect asthma severity, control, and healthcare utilization (28-30). While some studies have  
75 reported associations between particular air toxics and these asthma subphenotypes (9, 11,

76 12), none addressed our goal to identify combinations of air toxics from a large national  
77 assessment of air toxics associated with these asthma outcomes. We tested our hypothesis  
78 by geocoding levels of 125 air toxic from the US Environmental Protection Agency's (EPA)  
79 National Air Toxic Assessment (NATA) (31), one of the richest sources of multi-air toxic  
80 profiling across the US, to the residential addresses of children with asthma from our Airway  
81 in Asthma (ARIA) study (32) to map each child's exposure to air toxics during the first years  
82 of life. We addressed the challenges of combinatorial air toxic analysis by applying a  
83 machine learning-based algorithm called **Data-driven ExposurE Profile extraction (DEEP)**,  
84 which, to the best of our knowledge, is a novel method for this problem. DEEP uses the  
85 high-performing eXtreme Gradient Boosting (XGBoost) (33) algorithm to identify air toxic  
86 combinations associated with health outcomes. The combinations identified using XGBoost  
87 were then adjusted for potential confounders, including age, gender, race/ethnicity, and  
88 family income, to identify early-life multi-air toxic combinations, statistical interactions  
89 within combinations, and demographic profiles associated with adverse asthma outcomes in  
90 later childhood. Our approach identified several combinations of air toxics associated with  
91 asthma.

92

## 93 **Results**

94

## 95 **Characteristics of the study cohort**

96 **Table 1** shows the characteristics of the Airway in Asthma (ARIA) study (32) participants  
97 with asthma examined in this study. These 151 children with mild to severe persistent  
98 asthma were recruited from the Mount Sinai Health System, New York, NY with informed  
99 consent from their parent/guardian via an IRB-approved protocol. Participants had a mean  
100 age of 12 years (standard deviation 3.2 years) at the time of assessment, and were of  
101 diverse self-identified racial/ethnic backgrounds (**Table 1**). Their asthma was generally not  
102 well-controlled, with a mean score on the Asthma Control Test (ACT)(34) of 16.8 (maximum  
103 value 25 representing optimal control), and 96% of the cohort reporting regular use of a  
104 short-acting beta agonist rescue inhaler.

105 Children who used daily asthma controller medication (n=84, 56%) were younger than  
106 those who did not (n=65; p=0.048). Inhaled corticosteroids (ICS) were used most frequently,  
107 both independently and in combination with long-acting beta agonist (LABA). Children who  
108 had at least one lifetime emergency room visit for asthma (n=103, 68%) were more likely to  
109 self-identify as Black or Latino, had significantly lower ACT scores than their counterparts  
110 who had never required an emergency department visit for asthma, and were more likely to  
111 be taking combination ICS/LABA as their daily asthma controller medication. Children who  
112 had been hospitalized overnight for asthma in their lifetime (n=51, 34%) had significantly  
113 lower forced expiratory volume in 1 second percent predicted (FEV1%) on spirometry, and



114 higher rates of ICS/LABA and leukotriene receptor antagonist for daily asthma treatment  
115 compared to the participants with asthma who had never been hospitalized overnight for  
116 asthma.

117

### 118 **Air toxic characteristics**

119 Ambient annual average concentrations for over a hundred toxics based on emissions  
120 inventories and computer simulation models are publicly available for each US census tract  
121 in the EPA's NATA database (31). We mapped the available toxic levels to the residential ZIP  
122 code for each child in our cohort. Ninety-four zip codes spanning 443 square miles across  
123 New York, New Jersey, and Connecticut were represented in this cohort. We used the  
124 closest calendar year of NATA data available subsequent to a child's birth date. We retained  
125 only the air toxics whose levels were available for all the participants in the mapped  
126 datasets, yielding 125 air toxics for analysis.

127

### 128 ***DEEP-enabled identification of combinations of air toxics associated with childhood***

#### 129 ***asthma***

130 We then applied DEEP to identify air toxic combinations associated with each of the  
131 three childhood asthma outcomes, namely the need for daily asthma controller medication,  
132 lifetime emergency room visit for asthma, and lifetime overnight hospitalization for asthma.

133 In the first analytical stage of DEEP (detailed in Methods), for each outcome, the full dataset  
134 was randomly split 100 times into training and test sets in an 80:20 ratio. For each split, an  
135 XGBoost model consisting of 100 decision trees was learned from the training set and  
136 evaluated on the test set in terms of the Area Under the ROC Curve (AUC score) (35).

137 In the second analytical stage of DEEP, we analysed the combinations of toxics from the  
138 XGBoost models, identified as root-to-leaf paths in the constituent decision trees, for each  
139 outcome. Note that in some cases, a combination may consist of only one air toxic if it is  
140 sufficient to predict the outcome under consideration for a subset of the cohort, thus giving  
141 DEEP flexibility in discovery. Also, in cases of multiple air toxics in these combinations, their  
142 sequence of appearance on the path also indicates their relative order of relevance to the  
143 outcome being predicted. This is because variables closer to the root of a decision tree have  
144 higher predictive power than those closer to the leaves.

145 Next, the frequency of each combination was calculated as the number of models (out  
146 of 100) where it was included in at least one of the constituent trees. Candidate  
147 combinations were then identified as those with a frequency of at least ten. These  
148 combinations were then used in multivariable regression models to test their association  
149 with the asthma outcome of interest, while adjusting for age, gender, race/ethnicity, and  
150 income.

151           689 profiles of air toxics across all the asthma outcomes were discovered after the first  
152 XGBoost stage of DEEP. These sets included both individual air toxics and their combinations.  
153 359 of these sets were then found to be significantly associated ( $P \leq 0.05$ ) with the  
154 respective outcome in the second stage of DEEP. After multiple hypothesis correction by the  
155 Benjamini-Hochberg procedure (36), 273 air toxic profiles were found to be significantly  
156 associated ( $FDR \leq 0.05$ ) with at least one of the three outcomes. Our goal was to identify air  
157 toxic combinations whose increased levels are associated with adverse asthma outcomes.  
158 Therefore, among the significantly associated combinations, we focused on groups that  
159 included air toxics with levels higher than threshold. Among these finally determined  
160 combinations, there were 18 with only one air toxic each (**Figure 2**), and 20 multi-air toxic  
161 combinations (**Figure 3**).

162

### 163 ***Air toxic combinations associated with asthma outcomes***

164           Twenty multi-air toxic combinations and eighteen individual air toxics were found to be  
165 significantly associated with at least one of the three asthma outcomes. The medians and  
166 interquartile ranges of the exposure levels of the 34 air toxics included in these associations  
167 are shown in **Table 2**.

168           Higher levels of seventeen individual air toxics were significantly associated with worse  
169 asthma outcomes (**Figure 2**). Odds ratios (ORs) for these associations ranged from 1.56 to

170 2.65. Several of the identified toxics are established risk factors for childhood asthma,  
171 especially the chemicals previously categorized as halogenated, ketones and ethers (8,  
172 37-39). Among these, the air toxics most associated with the outcomes were acrylic acid  
173 (OR=2.10), mercury compounds (OR=2.65), and ethyl chloride (OR=1.87) respectively.  
174 Acetamide, pentachlorophenol, and polychlorinated biphenyls were associated with more  
175 than one asthma outcome.

176 A major strength of DEEP is its ability to identify multi-air toxic combinations associated  
177 with health outcomes. Indeed, here DEEP revealed significant associations between higher  
178 exposure to 20 multi-air toxic combinations and the three asthma outcomes of interest  
179 (**Figure 3**). Among these, 19 combinations included two air toxics and one included three.  
180 The associations of these combinations were generally stronger than those of the individual  
181 air toxics, with ORs ranging from 1.60 to 3.19 (**Figure 3**).

182 Notably, acrylic acid was not only the individual air toxic most strongly associated with  
183 daily controller medication (**Figure 2**), it was the first (i.e. primary) member of 7 of the 9  
184 multi-air toxic combinations associated with this outcome (**Figure 3**). Acrylic acid also  
185 appeared in 3 of the other 11 combinations associated with emergency room visit and  
186 overnight hospitalization for asthma (**Figure 3**), indicating that it is a major contributor to  
187 adverse asthma outcomes among children.

188 Three air toxic combinations were associated with lifetime emergency room visit for  
189 asthma, all with an OR of over 2 (**Figure 3**). Acetaldehyde, acrylamide, and acrylic acid were  
190 the primary exposures in these combinations, despite the fact they were not individually  
191 significantly associated with the outcome. Several other air toxics in these combinations,  
192 namely carbon disulfide and hydroquinone, were also not individually associated with this  
193 outcome. These findings highlight the main strength of DEEP, namely its ability to identify  
194 significant multi-air toxic combinations, whose constituent air toxics may not be individually  
195 associated with the health outcome of interest.

196 Among the eight air toxic combinations associated with lifetime overnight  
197 hospitalization for asthma, 1,4-dioxane, carbonyl sulfide, ethylidene dichloride, hydrochloric  
198 acid, and hydroquinone were the primary exposures (**Figure 3**). Both ethylidene dichloride  
199 and hydroquinone appeared in three of these eight combinations, indicating that these two  
200 chemicals may play a role in the development of poor asthma outcomes among children.  
201 Most other air toxics in these combinations (**Figure 3**) were largely not individually  
202 associated with this outcome (**Figure 2**), again supporting DEEP's ability to identify multi-air  
203 toxic combinations that may not be inferred from single air toxic associations.

204

205 **Effect sizes of multi-air toxic combinations may not be evident from the individual**  
206 **associations of their members**

207 Some air toxics had relatively low effect sizes when assessed individually (**Figure 2**)  
208 compared to the larger ORs from combination analyses (**Figure 3**). For example, acrylic acid  
209 was associated with daily controller medication with an OR of 2.10 as an individual air toxic  
210 (**Figure 2**), but the ORs of its combinations with dimethyl phthalate, 1,1,1-trichloroethane,  
211 ethyl chloride, acetophenone, and cobalt were higher (OR 2.16 to 3.19; **Figure 3**). Also, none  
212 of these five air toxics were individually associated with the outcome. Similarly,  
213 hexachlorobenzene was associated with daily controller medication with an OR of 2.03  
214 (**Figure 2**), while simultaneous exposure to the combination of hexachlorobenzene and  
215 dimethyl phthalate identified by DEEP had an OR of 2.96 (**Figure 3**). This was despite the fact  
216 that there was no significant individual association between dimethyl phthalate and the  
217 outcome. For the pair of toluene and phosphorus, neither air toxic was individually  
218 associated with daily controller medication (**Figure 2**), but their combination was associated  
219 with the outcome with an OR of 1.81 (**Figure 3**).

220 Similar cases of combinatorial effects were also seen for lifetime emergency room visit  
221 for asthma. For example, simultaneous exposure to polychlorinated biphenyl, acetaldehyde  
222 and carbon disulfide had 3.10-fold odds of the outcome (**Figure 3**), while polychlorinated  
223 biphenyl's individual effect size was substantially lower (OR=1.72; **Figure 2**). Similarly, the  
224 combination of acrylic acid and hydroquinone was significantly associated with emergency

225 room visit with an OR of 2.73 (**Figure 3**), but neither was associated with the outcome  
226 individually (**Figure 2**).

227 We observed similar results for multi-air toxic combinations and lifetime overnight  
228 hospitalization for asthma. Exposure to hydroquinone was individually associated with this  
229 outcome with an OR of 1.79 (**Figure 2**), but in combination with ethylidene dichloride, the  
230 association was stronger (OR=2.03; **Figure 3**). Similarly, carbonyl sulfide was not individually  
231 associated with this outcome (**Figure 2**), but it was the primary member in two of the  
232 multi-air toxic combinations found to be associated with OH (**Figure 3**).

233 In summary, the above comparison of the effect sizes of the individual (**Figure 2**) and  
234 multi- (**Figure 3**) air toxic associations demonstrated that combinations of air toxics had  
235 effects that were not fully explained by simply adding together the individual effects from  
236 their constituents. Overall, DEEP identified 34 air toxics associated with the asthma  
237 outcomes (**Table 2**), including 16 air toxics with significant effects only as members of  
238 combinations.

239

#### 240 **Statistical Interactions among Members of Air Toxic Combinations**

241 To assess potential synergy between members of air toxic combinations associated  
242 with asthma outcomes, we conducted statistical tests for interactions. Significant statistical  
243 interactions detected between air toxic members within the combinations are shown in

244 **Table 3.** Acrylic acid was the primary air toxic (i.e. primary branch point in the decision tree)  
245 of all the combinations with significant statistical interactions. While other combinations did  
246 not reveal significant interactions, such interactions remain possible given the limitations of  
247 statistical detection of interactions. Directed experimental work could be undertaken to test  
248 for additional interactions.

249

### 250 ***Representative Air Toxic Combinations and Demographic Risk Factors***

251 Finally, one of the advantages of DEEP is that the trees constituting its underlying  
252 XGBoost models can be visualized and interpreted, which is difficult to do for several other  
253 machine learning methods. However, since it is difficult to simultaneously depict all the  
254 trees inferred by DEEP, we visualized sample trees that contained the most strongly  
255 associated multi-air toxic combination for each childhood asthma outcome. Sample decision  
256 trees inferred by DEEP for each of the outcomes are shown in **Figures 4, 5, and 6**  
257 respectively. To provide an additional level of interpretation, we also compared the  
258 demographic characteristics (age, sex, race/ethnicity, and family income) of children  
259 exposed to each of these combinations to those of children who were not exposed in **Tables**  
260 **4, 5, and 6.** Differences could suggest demographic risk factors that may increase a child's  
261 exposure to these multi-air toxic combinations.



262 Acrylic acid and cobalt compounds was the air toxic combination associated with daily  
263 controller medication use with the highest odds ratio of 3.19 (**Figure 3, Figure 4**). Children  
264 exposed to this combination were older compared to those who were not ( $p=0.02$ ; **Table 4**).

265 Acetaldehyde, carbon disulphide, and polychlorinated biphenyls was the air toxic  
266 combination most strongly associated with lifetime emergency room visit for asthma  
267 ( $OR=3.10$ ; **Figure 3, Figure 5**). Children exposed to this combination were younger  
268 ( $p=5.34 \times 10^{-8}$ ; **Table 5**) and had lower family income than those who were not exposed  
269 ( $p=0.019$ ; **Table 5**). Exposed children were also less likely to be White ( $p=0.0046$ ; **Table 5**).  
270 These observations point to social disparities among these groups of children.

271 The most strongly associated combination for overnight hospitalization was  
272 hydroquinone and ethylidene dichloride ( $OR=2.03$ ; **Figure 3, Figure 6**). Children exposed to  
273 this combination were younger ( $p=0.00218$ ; **Table 6**) and had lower family incomes  
274 ( $p=8.26 \times 10^{-5}$ ; **Table 6**) than those who were not exposed.

275

## 276 **Discussion**

277 Our application of a novel machine learning-driven algorithm called DEEP to a cohort of  
278 mild to severe asthmatic children identified several individual and combinations of air toxics,  
279 to which increased exposure during early-life was associated with adverse asthma outcomes  
280 in later childhood. In particular, due to a unique ability of DEEP to examine air toxic

281 combinations, we identified 16 air toxics that were only found to be significantly associated  
282 with childhood asthma outcomes in combination with other air toxics.

283 Many air toxics in the identified combinations, such as carbonyl sulfide, carbon  
284 disulfide, ethyl chloride, and ethylidene chloride, are similar in structure and have analogous  
285 formation, production, chemical fate, and chemical transport properties (40). Ten air toxics  
286 in the combinations contained chlorine, three included heavy metal compounds, and many  
287 were acidic chemicals. This aligns with prior literature implicating acidic chemicals,  
288 chlorinated chemicals, and heavy metal compounds as risk factors for asthma and asthma  
289 severity (8, 41-44). However, an understanding of the biological mechanisms through which  
290 these combinations of air toxics can jointly affect respiratory health and asthma merits  
291 further study.

292 Among the air toxics individually associated with asthma outcomes (**Figure 2**),  
293 triethylamine was associated with increased overnight hospitalizations for asthma.  
294 Triethylamine is a clear, colorless liquid used in waterproofing and as a catalyst, corrosion  
295 inhibitor, and propellant (45). It is a respiratory irritant, to which chronic exposure even at  
296 low levels can inhibit the function of organic cationic transporters, thus preventing efficient  
297 uptake of inhaled bronchodilators used to control acute asthma symptoms (46, 47).

298 Acrylic acid was individually associated with daily controller medication (**Figure 2**) and  
299 appeared as a member of at least one combination associated with all three outcomes

300 **(Figure 3)**. Furthermore, it was found to interact with other member air toxics of three  
301 combinations **(Table 3)**. Acrylic acid is used in the manufacture of adhesives, elastomers,  
302 plastics, coatings, as well as floor paints and polishers (48). Literature has suggested that the  
303 presence of water-soluble cobalt complexes increases the conversion of polyacrylic acid into  
304 acrylic acid, which is more biologically viable. Acrylic acid also reacts with cobalt complexes  
305 to produce organocobalt complexes (49). Additionally, hydroquinone acts as a stabilizer to  
306 prevent the polymerization of acrylic acid, which keeps the latter in a form with a lower  
307 molecular weight that is more biologically viable (50, 51). Our results, including evidence of  
308 statistical interactions between acrylic acid and other chemical compounds, suggest further  
309 investigation of mechanisms for acrylic acid's associations with adverse childhood asthma  
310 outcomes.

311 Ethyl chloride, also known as chloroethane ( $C_2H_5Cl$ ), and ethylidene dichloride ( $C_2H_4Cl_2$ )  
312 are both chlorinated hydrocarbons. Ethyl chloride is used as a thickening agent and binder in  
313 paints and cosmetics, and as a refrigerant, aerosol spray propellant, anesthetic, and blowing  
314 agent for foam packaging (52). We found ethyl chloride to be associated with asthma  
315 outcomes, both as an individual air toxic (overnight hospitalizations; **Figure 2**) and as a  
316 member of multi-air toxic combinations (daily controller medication; **Figure 3**). Ethylidene  
317 dichloride, which is mainly used as a solvent for plastics, oils and fats, and as a degreaser  
318 and fumigant in insecticide sprays (53), appeared as a member of several combinations

319 associated with lifetime overnight hospitalization for asthma (**Figure 3**). Both these  
320 compounds are well-known members of the chloroethanes family, which is comprised of  
321 liposoluble chemicals that can be taken up by the lipoprotein within the alveolar film layer  
322 (AFL) (54). AFL disruption is observed in multiple pulmonary diseases, including acute  
323 respiratory distress syndrome, infant respiratory distress syndrome, emphysema, chronic  
324 obstructive pulmonary disease, asthma, chronic bronchitis, pneumonia, pulmonary  
325 infections, and idiopathic pulmonary fibrosis (55). Thus, chronic exposure to ethyl chloride  
326 and ethylidene dichloride may lead to dysfunction in the AFL, which may contribute to  
327 worse asthma control.

328       Hydroquinone, a commonly studied air toxic, was also identified in our analyses. While  
329 exposure to higher levels of hydroquinone alone was not associated with overnight  
330 hospitalization for asthma, DEEP identified it as a member of several multi-air toxic  
331 combinations associated with this outcome (**Figure 3**). Hydroquinone is commonly found in  
332 the indoor environment, and exposure to it has been associated with airway  
333 hypersensitivity (56-59). Hydroquinone is widely seen in cosmetic and health products,  
334 including skin creams (60). It is thought to prevent the polymerization of acrylic acid, methyl  
335 methacrylate, cyanoacrylate, and other monomers that are susceptible to radical-initiated  
336 polymerization, thus allowing them to persist in their original form (50, 51). This suggests a  
337 mechanism through which the identified synergistic combination of hydroquinone with

338 acrylic acid (emergency room visit; **Figure 3**) is associated with adverse asthma outcomes.  
339 Although our analysis did not find a statistically significant interaction between  
340 hydroquinone and ethylidene dichloride, potentially due to lower exposure levels,  
341 hydroquinone is industrially added to shelf ethylidene dichloride as a stabilizer (61). This  
342 suggests that in the presence of hydroquinone, ethylidene dichloride is less likely to react  
343 with other chemicals in the environment, and thus retain its toxic form longer (similar to  
344 acrylic acid). Thus, it is still possible that hydroquinone and ethylidene dichloride may act  
345 synergistically, but this needs to be investigated in future studies.

346 While our study has advanced the identification of air toxic combinations associated  
347 with childhood asthma outcomes, it also has limitations. We used the NATA national model  
348 (31) to estimate exposures rather than personal sampling or local monitors. Collecting  
349 personal or locally monitored measures for 125 air toxics at each cohort participant's  
350 residence would be logistically and financially challenging. Given this, NATA is commonly  
351 used for estimating ambient exposures, since it is a well-validated deterministic dispersion  
352 chemical transportation model created by the EPA that accounts for sources included in the  
353 NATA emission inventory (31). NATA estimates of a given air toxic may under-report a  
354 personally or locally monitored value, since the latter may include emissions from indoor  
355 and undocumented sources not in the EPA's inventory. For instance, one study found that  
356 higher personally monitored benzene concentrations relative to NATA-predicted values

357 were likely due, at least in part, to indoor sources not included in the EPA's inventory (62).  
358 Other studies also found discrepancies between NATA estimates and monitored chemical  
359 concentrations, due again, in part, to local or indoor sources (63, 64).

360 Despite the above limitations, using NATA as the primary source of exposure estimates  
361 has several strengths over locally monitored values. First, NATA has a finer geographical  
362 prediction resolution and spread than currently available monitoring sites (31). This enabled  
363 us to include participants in our study that may not have had a monitoring site close to their  
364 residence. Second, NATA data are generated from an advanced chemical transportation  
365 model that aggregates exposure over a long period, and thus is able to capture transient  
366 exposures. Also, several factors potentially affecting air toxic estimates, such as seasonality,  
367 ambient temperature, meteorology, precipitation, and solar radiation, have already been  
368 incorporated into NATA's model (31). This level of comprehensive modelling is typically not  
369 available from personal or local monitoring. Finally, local measurements may also have  
370 detection and quantification limits, while NATA is able to estimate air toxics even at lower  
371 levels and over a longer period of time. Due to these strengths, we chose to study NATA  
372 data in this work.

373 We recognize that our results do not provide evidence of a causal effect of any  
374 chemical on adverse childhood asthma outcomes, which will need further investigation.  
375 Additionally, our geocoding was based on a single zip code for each participant so would not

376 account for potentially dynamic exposures due to residential moves. Last, our study  
377 participants were all from the New York, New Jersey, and Connecticut tri-state area. Thus,  
378 our findings may not generalize to other US regions or parts of the world. Future studies  
379 could examine combinations from other geographical regions and/or utilize direct air  
380 sampling to confirm the combinations identified in this study.

381 In conclusion, this study demonstrated innovative use of data science methods and  
382 data sources to identify specific combinations of early-life air toxic exposures associated  
383 with later childhood asthma outcomes. Our study suggests that chemical pollutants should  
384 be closely monitored together in combination, especially for locations with vulnerable  
385 populations.

386

## 387 **Materials and Methods**

388 An overview of the study approach is shown in **Figure 1**.

389

### 390 ***Study Population***

391 The study population included children with asthma from the AiRway In Asthma (ARIA)  
392 study, a cohort recruited from the Mount Sinai Health System, New York, NY (32). The study  
393 was approved by the Mount Sinai Institutional Review Board. Children with asthma had mild  
394 to severe persistent asthma according to National Asthma Education and Prevention

395 Program (NAEPP)/National Heart Lung Blood Institute (NHLBI) Expert Panel Report 3 (EPR3)  
396 criteria (65) and positive bronchodilator response on spirometry or methacholine challenge  
397 with provocative challenge causing a 20% fall in forced expiratory volume (PC20) < 12.5  
398 mg/ml. Phenotyping for all participants included detailed questionnaires about  
399 asthma-related symptoms, medication and healthcare use, and pre and post-  
400 bronchodilator spirometry following American Thoracic Society guidelines (65).

401

#### 402 ***Asthma Outcomes***

403 We focused on the following three self-reported asthma-related outcomes: (1) current use  
404 of prescribed daily asthma controller medication (daily controller medication), (2) at least  
405 one emergency department visit for asthma during subject's lifetime (emergency room visit),  
406 and (3) at least one overnight hospitalization for asthma during subject's lifetime (overnight  
407 hospitalization).

408

#### 409 ***Air Toxic Exposures***

410 The air toxic exposure profile of each participant was derived from EPA's National Air Toxics  
411 Assessment (NATA) (31). NATA estimates the annual ambient concentrations of over a  
412 hundred air toxics at each census tract in the United States based on emissions inventories  
413 and advanced computer simulation models (31). Seasonality of air toxics, ambient



414 temperature, meteorology, precipitation and solar radiation are incorporated into NATA's  
415 model (31).

416 NATA data are available for 1996, 1999, 2002, 2005, 2011, and 2014 (27), and children  
417 in this study were born between 1997 and 2012. To assign the most representative ambient  
418 air toxic exposure levels to each participant, we mapped the residential ZIP code of each  
419 child in our cohort to the geometric centroid of the closest census tract (31). We then used  
420 the annual exposure data for that tract from the NATA release closest in time following the  
421 child's birth year. This choice of year closest to birth was based on prior evidence that  
422 early-life exposure to air pollution is associated with childhood asthma outcomes (66-68).  
423 Finally, we retained the 125 air toxics that had data available, i.e., no missing data, for all  
424 participants in the final dataset (15, 69).

425

#### 426 ***Covariates***

427 We included age, sex, race/ethnicity, and family income as covariates in multivariable  
428 regression models based on considerations that these variables could confound associations  
429 between air toxic levels and asthma outcomes. Since the questionnaire completed by ARIA  
430 participants did not include queries about family income, we used the average income of  
431 each participant's residential ZIP code obtained from US Census Business Patterns data (70)  
432 as a surrogate for this variable.

433

434 ***Data-driven Exposure Profile extraction (DEEP)***

435 To identify multi-air toxic exposure profiles associated with asthma outcomes, we  
436 developed a data-driven method called Data-driven Exposure Profile extraction (DEEP)  
437 (**Figure 1**). DEEP is inspired by a simpler method that we previously used to identify multi-air  
438 toxic combinations associated with children's cognitive skills (15).

439 In the first stage of DEEP, exposure combinations are identified using eXtreme Gradient  
440 Boosting (XGBoost) (33), an algorithm that uses an ensemble method to iteratively learn  
441 decision trees and generally performs well at prediction tasks. XGBoost generally yields  
442 strong predictive power (71, 72) due to its use of multiple optimization methods, including  
443 regularization and gradient boosting, which reduces overfitting of models to training data.  
444 Specifically, the full exposure dataset was randomly split 100 times into training and test  
445 sets in an 80:20 ratio. For each split, an XGBoost model consisting of 100 decision trees was  
446 learned from the training set to predict the outcome under consideration. This model was  
447 then applied to and evaluated on the corresponding test set in terms of the Area Under the  
448 ROC curve (AUC score(35)). The overall predictability of the target outcome was evaluated  
449 in terms of the average value of the AUC scores across the 100 training/test splits.

450 The decision trees constituting each XGBoost model contain internal decision nodes,  
451 edges, and leaf nodes to represent how the value of an outcome could be predicted based

452 on air toxic levels. **Figures 4-6** show several trees derived in the current work. Each decision  
453 node in these trees contains an air toxic and a threshold value for its level. It is also  
454 connected by two edges representing the decisions made depending on whether an  
455 individual's exposure was higher or lower than the threshold. Each of these edges is  
456 connected to either the next decision node or a leaf node. A leaf node determines the value  
457 of the outcome for the individual with the exposure profile represented by the decision path  
458 taken to reach it. Each decision and leaf node also represents a subpopulation of the cohort  
459 exposed to the air toxics on the path taken to reach it. Candidate multi-air toxic  
460 combinations are then defined as the air toxics and thresholds in the decision nodes  
461 constituting the paths from the root of a tree to the leaf nodes. We calculated the frequency  
462 of each combination as the number of XGBoost models (out of 100) where it was included in  
463 at least one of the constituent trees, and set of threshold of 10 to identify the most relevant  
464 combinations. Note that, if two or more variables are highly correlated, and thus similarly  
465 associated with the outcome, a key characteristic of the decision trees in the XGBoost  
466 model is that they will include only one of these variables as an internal decision node. Thus,  
467 unlike traditional regression models, XGBoost is not as adversely affected by collinearity  
468 among the input variables. Furthermore, DEEP executes XGBoost 100 times on randomly  
469 selected training sets, different selections of these variables may be included in the different  
470 trees inferred, thus enhancing the coverage of the air toxic profiles.

471 In the second stage of DEEP, a multivariable linear regression model is built to assess  
472 the association of a candidate combination with the target outcome, adjusted for covariates.  
473 The asthma outcome is the dependent variable in this model, while the air toxic  
474 combination and covariates are its independent variables. The variable representing the  
475 combination takes a value of 1 for individuals exposed to it, determined using the threshold  
476 values of the constituent air toxics, and 0 otherwise. One model is built for each outcome  
477 and candidate combination, yielding the odds ratio (OR) denoting the strength of the  
478 association between the two. The p-values of all the associations are converted into false  
479 discovery rates (FDRs) after correcting for multiple hypothesis testing using the  
480 Benjamini-Hochberg method (36). In this study, significant associations were identified as  
481 those with  $FDR \leq 0.05$ .

482 To assess potential synergy between members of air toxic combinations associated  
483 with asthma outcomes, we conducted statistical tests for interactions. Interactions between  
484 pairs of air toxics were assessed through additional multivariable regression models where  
485 the outcome was the dependent variable and predictors included the levels of the two  
486 toxics, their product as a representative of their interaction, and covariates. For  
487 combinations with two air toxics, this regression model was inferred from the whole cohort,  
488 while for combinations with three air toxics, analyses were conducted for the last two toxics  
489 on the sample meeting the threshold for the first toxic level in the combination. A significant

490 interaction was identified if the p-value of the interaction term in the model was lower than  
491 0.05.

492 The DEEP framework is implemented in the Python programming language (71). The  
493 XGBoost, model evaluation (AUC score calculation) and regression components are  
494 implemented using the xgboost (33), scikit-learn (72) and statsmodels (73) Python packages  
495 respectively.

496

#### 497 **Study Approval**

498 The study was approved by the Mount Sinai Institutional Review Board, New York, USA.

499 Parents of participants provided written informed consent.

500

#### 501 **Author contributions**

502 HLH, PC, GP and SB conceived the study. GP and SB supervised the work. YCL, HLH, PC, ZA,

503 and YC managed and analyzed the data. YCL and HLH drafted the manuscript. All the

504 authors reviewed, edited, and approved the manuscript. Order among co-first authors was

505 determined based on contribution to results generation.

506

#### 507 **Acknowledgements**

508 This work was supported by a pilot grant from the Department of Genetics and Genomic  
509 Sciences at Mount Sinai and NIH grants R01 HL147328, R01 AI118833, UG3 OD023337,  
510 R01HG011407-01A1, and P30 ES23515. It was also supported in part through the  
511 computational resources provided by Scientific Computing at the Icahn School of Medicine  
512 at Mount Sinai. We thank Alfin Vicencio of the Mount Sinai Hospital System for his  
513 assistance with cohort recruitment and Jeanette Stingone of Columbia University for her  
514 technical advice.

515

516 **Conflicts of interest**

517 The authors have declared that no conflict of interest exists.

518 **References**

- 519 1. United States Environmental Protection Agency. About Urban Air Toxics  
520 <https://www.epa.gov/urban-air-toxics/about-urban-air-toxics>. Accessed 8/12/2021.
- 521 2. Walters GI, Robertson AS, Moore VC, and Burge PS. Occupational asthma caused by  
522 acrylic compounds from SHIELD surveillance (1989-2014). *Occup Med (Lond)*.  
523 2017;67(4):282-9.
- 524 3. North ML, Takaro TK, Diamond ML, and Ellis AK. Effects of phthalates on the  
525 development and expression of allergic disease and asthma. *Annals of allergy, asthma*  
526 *& immunology : official publication of the American College of Allergy, Asthma, &*  
527 *Immunology*. 2014;112(6):496-502.
- 528 4. Dumas O, Despreaux T, Perros F, Lau E, Andujar P, Humbert M, et al. Respiratory  
529 effects of trichloroethylene. *Respiratory medicine*. 2018;134:47-53.
- 530 5. Schenker MB, and Jacobs JA. Respiratory effects of organic solvent exposure.  
531 *Tubercle and lung disease : the official journal of the International Union against*  
532 *Tuberculosis and Lung Disease*. 1996;77(1):4-18.
- 533 6. Nurmatov UB, Tagiyeva N, Semple S, Devereux G, and Sheikh A. Volatile organic  
534 compounds and risk of asthma and allergy: a systematic review. *European respiratory*  
535 *review : an official journal of the European Respiratory Society*.  
536 2015;24(135):92-101.
- 537 7. Rumchev K, Spickett J, Bulsara M, Phillips M, and Stick S. Association of domestic  
538 exposure to volatile organic compounds with asthma in young children. *Thorax*.  
539 2004;59(9):746-51.
- 540 8. Weisel CP. Assessing exposure to air toxics relative to asthma. *Environ Health*  
541 *Perspect*. 2002;110 Suppl 4:527-37.
- 542 9. Ye D, Klein M, Chang HH, Sarnat JA, Mulholland JA, Edgerton ES, et al. Estimating  
543 Acute Cardiorespiratory Effects of Ambient Volatile Organic Compounds.  
544 *Epidemiology*. 2017;28(2):197-206.
- 545 10. Rosa MJ, Jung KH, Perzanowski MS, Kelvin EA, Darling KW, Camann DE, et al.  
546 Prenatal exposure to polycyclic aromatic hydrocarbons, environmental tobacco smoke  
547 and asthma. *Respiratory medicine*. 2011;105(6):869-76.
- 548 11. Ran J, Kioumourtzoglou MA, Sun S, Han L, Zhao S, Zhu W, et al. Source-Specific  
549 Volatile Organic Compounds and Emergency Hospital Admissions for  
550 Cardiorespiratory Diseases. *International journal of environmental research and*  
551 *public health*. 2020;17(17).
- 552 12. Delfino RJ. Epidemiologic evidence for asthma and exposure to air toxics: linkages  
553 between occupational, indoor, and community air pollution research. *Environmental*  
554 *health perspectives*. 2002;110 Suppl 4:573-89.

- 555 13. Bably M, Arif AA, and Post A. Prenatal use of cleaning and scented products and its  
556 association with childhood asthma, asthma symptoms, and mental health and  
557 developmental comorbidities. *The Journal of asthma : official journal of the*  
558 *Association for the Care of Asthma*. 2021;58(1):46-51.
- 559 14. Adgent MA, Carroll KN, Hazlehurst MF, Loftus CT, Szpiro AA, Karr CJ, et al. A  
560 combined cohort analysis of prenatal exposure to phthalate mixtures and childhood  
561 asthma. *Environment international*. 2020;143:105970.
- 562 15. Stingone JA, Pandey OP, Claudio L, and Pandey G. Using machine learning to  
563 identify air pollution exposure profiles associated with early cognitive skills among  
564 U.S. children. *Environ Pollut*. 2017;230:730-40.
- 565 16. Patel CJ. Analytic Complexity and Challenges in Identifying Mixtures of Exposures  
566 Associated with Phenotypes in the Exposome Era. *Curr Epidemiol Rep*.  
567 2017;4(1):22-30.
- 568 17. Choi H, Schmidbauer N, Sundell J, Hasselgren M, Spengler J, and Bornehag CG.  
569 Common household chemicals and the allergy risks in pre-school age children. *PLoS*  
570 *one*. 2010;5(10):e13423.
- 571 18. Oakes M, Baxter L, and Long TC. Evaluating the application of multipollutant  
572 exposure metrics in air pollution health studies. *Environ Int*. 2014;69:90-9.
- 573 19. Khreis H, Kelly C, Tate J, Parslow R, Lucas K, and Nieuwenhuijsen M. Exposure to  
574 traffic-related air pollution and risk of development of childhood asthma: A  
575 systematic review and meta-analysis. *Environ Int*. 2017;100:1-31.
- 576 20. Esposito S, Tenconi R, Lelii M, Preti V, Nazzari E, Consolo S, et al. Possible  
577 molecular mechanisms linking air pollution and asthma in children. *BMC Pulm Med*.  
578 2014;14:31.
- 579 21. Orellano P, Quaranta N, Reynoso J, Balbi B, and Vasquez J. Effect of outdoor air  
580 pollution on asthma exacerbations in children and adults: Systematic review and  
581 multilevel meta-analysis. *PLoS One*. 2017;12(3):e0174050.
- 582 22. Edwards SC, Jedrychowski W, Butscher M, Camann D, Kieltyka A, Mroz E, et al.  
583 Prenatal exposure to airborne polycyclic aromatic hydrocarbons and children's  
584 intelligence at 5 years of age in a prospective cohort study in Poland. *Environ Health*  
585 *Perspect*. 2010;118(9):1326-31.
- 586 23. Gehring U, Wijga AH, Hoek G, Bellander T, Berdel D, Bruske I, et al. Exposure to  
587 air pollution and development of asthma and rhinoconjunctivitis throughout childhood  
588 and adolescence: a population-based birth cohort study. *Lancet Respir Med*.  
589 2015;3(12):933-42.
- 590 24. Chang HH, Pan A, Lary DJ, Waller LA, Zhang L, Brackin BT, et al. Time-series  
591 analysis of satellite-derived fine particulate matter pollution and asthma morbidity in  
592 Jackson, MS. *Environ Monit Assess*. 2019;191(Suppl 2):280.

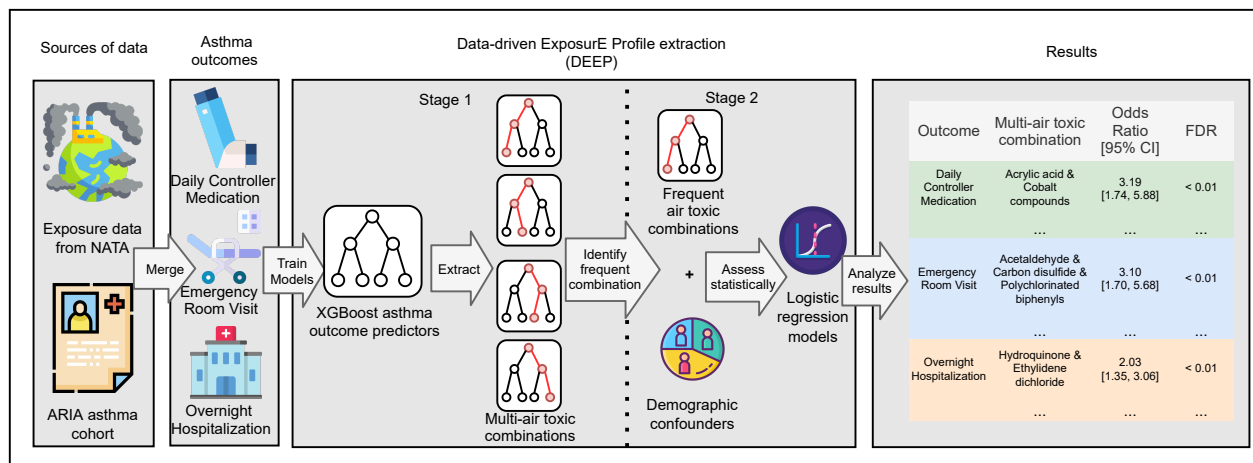


- 593 25. Brokamp C, Jandarov R, Rao MB, LeMasters G, and Ryan P. Exposure assessment  
594 models for elemental components of particulate matter in an urban environment: A  
595 comparison of regression and random forest approaches. *Atmos Environ (1994)*.  
596 2017;151:1-11.
- 597 26. Coker E, Liverani S, Ghosh JK, Jerrett M, Beckerman B, Li A, et al. Multi-pollutant  
598 exposure profiles associated with term low birth weight in Los Angeles County.  
599 *Environ Int*. 2016;91:1-13.
- 600 27. Deng H, Urman R, Gilliland FD, and Eckel SP. Understanding the importance of key  
601 risk factors in predicting chronic bronchitic symptoms using a machine learning  
602 approach. *BMC Med Res Methodol*. 2019;19(1):70.
- 603 28. Centers for Disease Control and Prevention. Measures to Identify and Track Racial  
604 Disparities in Childhood Asthma: Prevalence and Outcome Measures. July 14, 2016.  
605 Available at  
606 [https://www.cdc.gov/asthma/asthma\\_disparities/outcome\\_measures.htm#anchor\\_153](https://www.cdc.gov/asthma/asthma_disparities/outcome_measures.htm#anchor_1532268838154)  
607 [2268838154](https://www.cdc.gov/asthma/asthma_disparities/outcome_measures.htm#anchor_1532268838154), accessed 8/10/2021.
- 608 29. Gliklich RE, Castro M, Leavy MB, Press VG, Barochia A, Carroll CL, et al.  
609 Harmonized outcome measures for use in asthma patient registries and clinical  
610 practice. *The Journal of allergy and clinical immunology*. 2019;144(3):671-81 e1.
- 611 30. Covar RA, Fuhlbrigge AL, Williams P, and Kelly HW. The Childhood Asthma  
612 Management Program (CAMP): Contributions to the Understanding of Therapy and  
613 the Natural History of Childhood Asthma. *Current respiratory care reports*.  
614 2012;1(4):243-50.
- 615 31. United States Environmental Protection Agency. National-Scale Air Toxics  
616 Assessment. Washington, DC: U.S. Environmental Protection Agency. 2014.  
617 Available at <https://www.epa.gov/national-air-toxics-assessment>, accessed 12/10/2020.
- 618 32. Do AN, Chun Y, Grishina G, Grishin A, Rogers AJ, Raby BA, et al. Network study of  
619 nasal transcriptome profiles reveals master regulator genes of asthma. *J Allergy Clin*  
620 *Immunol*. 2021;147(3):879-93.
- 621 33. Tianqi Chen CG. XGBoost: A Scalable Tree Boosting System. *In Proceedings of the*  
622 *22nd ACM SIGKDD International Conference on Knowledge Discovery and Data*  
623 *Mining Association for Computing Machinery, New York, NY, USA*. 2016:785–94. .
- 624 34. Nathan RA, Sorkness CA, Kosinski M, Schatz M, Li JT, Marcus P, et al.  
625 Development of the asthma control test: a survey for assessing asthma control. *The*  
626 *Journal of allergy and clinical immunology*. 2004;113(1):59-65.
- 627 35. Lever J, Krzywinski M, and Altman N. Classification evaluation. *Nature Methods*.  
628 2016;13(8):603-4.

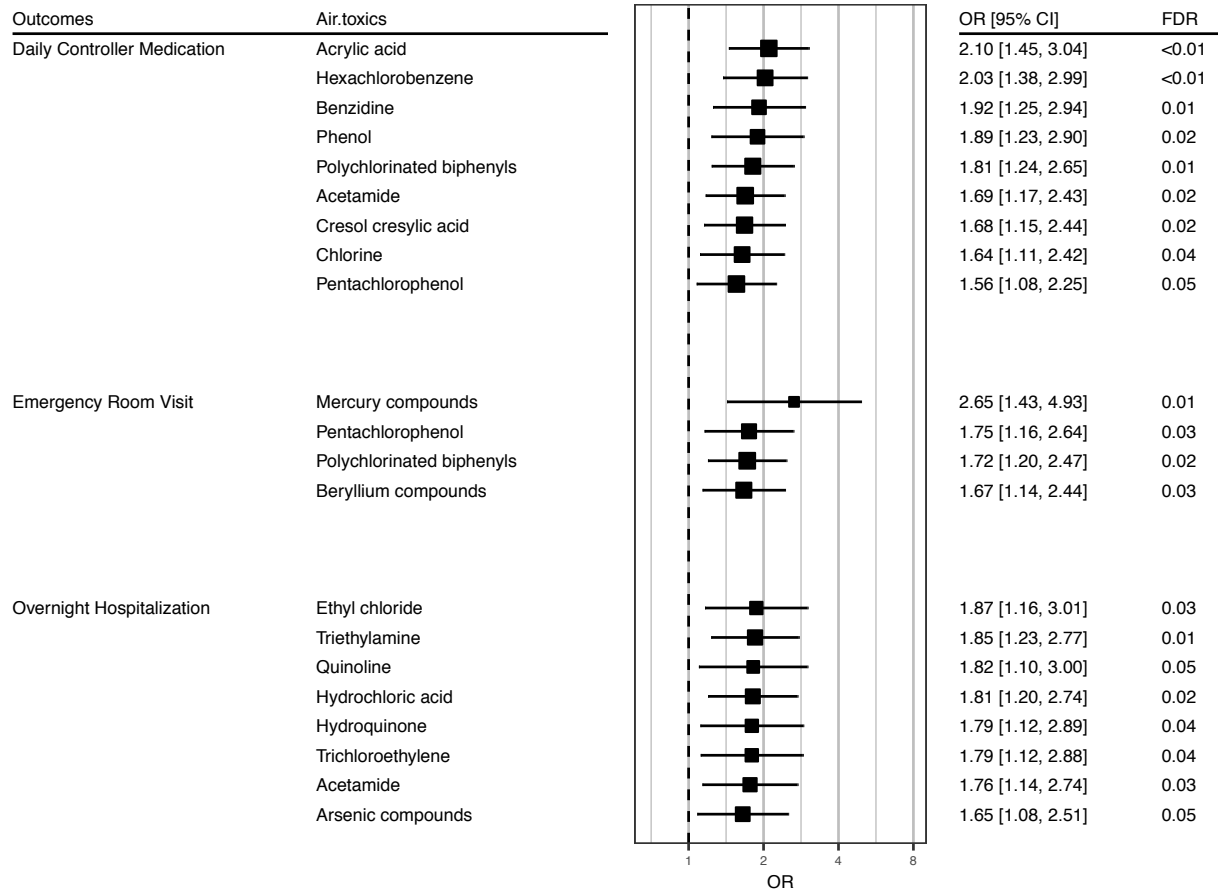
- 629 36. Benjamini Y, and Hochberg Y. Controlling the False Discovery Rate: A Practical and  
630 Powerful Approach to Multiple Testing. *Journal of the Royal Statistical Society Series*  
631 *B (Methodological)*. 1995;57(1):289-300.
- 632 37. Chin JY, Godwin C, Parker E, Robins T, Lewis T, Harbin P, et al. Levels and sources  
633 of volatile organic compounds in homes of children with asthma. *Indoor air*.  
634 2014;24(4):403-15.
- 635 38. Gahleitner F, Guallar-Hoyas C, Beardsmore CS, Pandya HC, and Thomas CP.  
636 Metabolomics pilot study to identify volatile organic compound markers of childhood  
637 asthma in exhaled breath. *Bioanalysis*. 2013;5(18):2239-47.
- 638 39. Delfino RJ, Gong H, Linn WS, Hu Y, and Pellizzari ED. Respiratory symptoms and  
639 peak expiratory flow in children with asthma in relation to volatile organic  
640 compounds in exhaled breath and ambient air. *J Expo Anal Environ Epidemiol*.  
641 2003;13(5):348-63.
- 642 40. Ferm RJ. The Chemistry Of Carbonyl Sulfide. *Chemical Reviews*. 1957;57(4):621-40.
- 643 41. Moghtaderi M, Ashraf MA, Moghtaderi T, Teshnizi SH, and Nabavizadeh SH. Heavy  
644 metal concentration in classroom dust samples and its relationship with childhood  
645 asthma: a study from Islamic Republic of Iran. *Eastern Mediterranean health journal*  
646 *= La revue de sante de la Mediterranee orientale = al-Majallah al-sihhiyah li-sharq*  
647 *al-mutawassit*. 2020;26(5):594-601.
- 648 42. Carder M, Seed MJ, Money A, Agius RM, and van Tongeren M. Occupational and  
649 work-related respiratory disease attributed to cleaning products. *Occup Environ Med*.  
650 2019;76(8):530-6.
- 651 43. Hsieh CY, Jung CR, Lin CY, and Hwang BF. Combined exposure to heavy metals in  
652 PM2.5 and pediatric asthma. *J Allergy Clin Immunol*. 2020.
- 653 44. Andersson M, Backman H, Nordberg G, Hagenbjork A, Hedman L, Eriksson K, et al.  
654 Early life swimming pool exposure and asthma onset in children - a case-control  
655 study. *Environ Health*. 2018;17(1):34.
- 656 45. New Jersey Department of Health. Hazardous Substance Fact Sheet: Triethylamine.  
657 2010. Available at <https://nj.gov/health/eoh/rtkweb/documents/fs/1907.pdf>, accessed  
658 8/16/2021.
- 659 46. Mukherjee M, Cingolani E, Pritchard DI, and Bosquillon C. Enhanced expression of  
660 Organic Cation Transporters in bronchial epithelial cell layers following insults  
661 associated with asthma - Impact on salbutamol transport. *Eur J Pharm Sci*.  
662 2017;106:62-70.
- 663 47. Laborde-Casterot H, Rosenberg N, Dupont P, and Garnier R. Is the incidence of  
664 aliphatic amine-induced occupational rhinitis and asthma underestimated? *Am J Ind*  
665 *Med*. 2014;57(12):1303-10.

- 666 48. Ohara T, Sato T, Shimizu N, Prescher G, Schwind H, Weiberg O, et al. *Ullmann's*  
667 *Encyclopedia of Industrial Chemistry*. 2020:1-21.
- 668 49. Peng C-H, Fryd M, and Wayland BB. Organocobalt Mediated Radical Polymerization  
669 of Acrylic Acid in Water. *Macromolecules*. 2007;40(19):6814-9.
- 670 50. Wang X, Schmidt F, Hanaor D, Kamm PH, Li S, and Gurlo A. Additive  
671 Manufacturing of Ceramics from Pre-ceramic Polymers: A Versatile  
672 Stereolithographic Approach Assisted by Thiol-Ene Click Chemistry. *Additive*  
673 *Manufacturing*. 2019;27.
- 674 51. Pozdeeva NN, and Denisov ET. Mechanism of hydroquinone-inhibited oxidation of  
675 acrylic acid and methyl methacrylate.
- 676 52. Rossberg M, Lendle W, Pfliederer G, Tögel A, Dreher E-L, Langer E, et al.  
677 *Ullmann's Encyclopedia of Industrial Chemistry*. 2006.
- 678 53. Agency for Toxic Substances and Disease Registry. Toxicological Profile for  
679 1,1-Dichloroethane. U.S. Department of Health and Human Services, Atlanta, GA.;  
680 1990.
- 681 54. Fiserova-Bergerova V, Tichy M, and Di Carlo FJ. Effects of biosolubility on  
682 pulmonary uptake and disposition of gases and vapors of lipophilic chemicals. *Drug*  
683 *Metab Rev*. 1984;15(5-6):1033-70.
- 684 55. Holder JW. Physical and physicochemical factors effecting transport of  
685 chlorohydrocarbon gases from lung alveolar air to blood as measured by the causation  
686 of narcosis. *J Environ Sci Health C Environ Carcinog Ecotoxicol Rev*.  
687 2012;30(1):42-80.
- 688 56. Becher R, Hongslo JK, Jantunen MJ, and Dybing E. Environmental chemicals  
689 relevant for respiratory hypersensitivity: the indoor environment. *Toxicol Lett*.  
690 1996;86(2-3):155-62.
- 691 57. Sivertsen B, and Clench-Aas J. Exposure to environmental chemicals relevant for  
692 respiratory hypersensitivity: European aspects. *Toxicol Lett*. 1996;86(2-3):143-53.
- 693 58. Schwela D. Exposure to environmental chemicals relevant for respiratory  
694 hypersensitivity: global aspects. *Toxicol Lett*. 1996;86(2-3):131-42.
- 695 59. Triggiani M, Loffredo S, Granata F, Staiano RI, and Marone G. Modulation of mast  
696 cell and basophil functions by benzene metabolites. *Curr Pharm Des*.  
697 2011;17(34):3830-5.
- 698 60. Olumide YM, Akinkugbe AO, Altraide D, Mohammed T, Ahamefule N, Ayanlowo S,  
699 et al. Complications of chronic use of skin lightening cosmetics. *International Journal*  
700 *of Dermatology*. 2008;47(4):344-53.
- 701 61. World Health Organization, United Nations Environment Programme on Chemical  
702 Safety, International Labour Organisation. Hydroquinone: Health and Safety Guide.  
703 Geneva: World Health Organization, 1996.

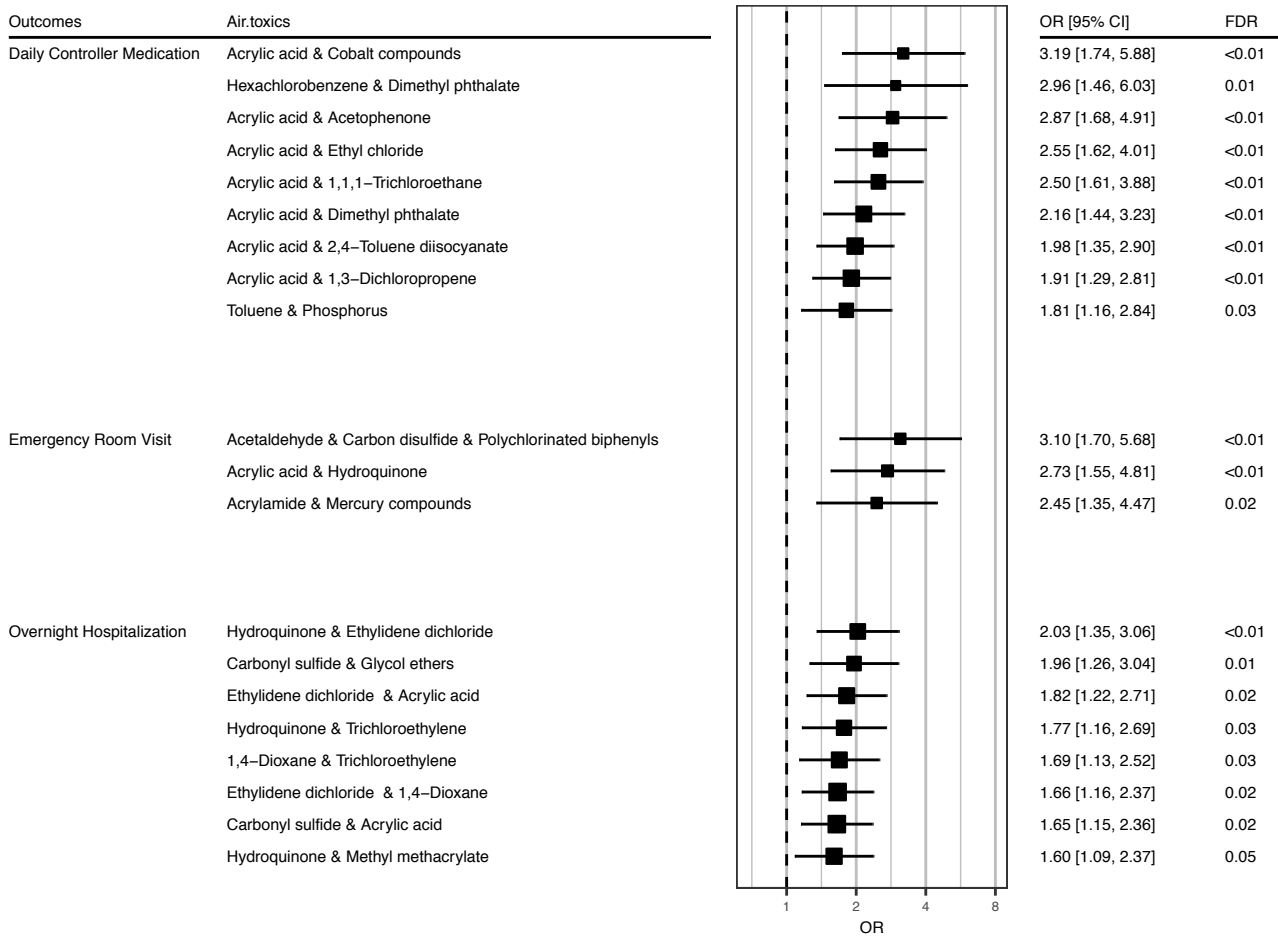
- 704 62. George B, Schultz B, Palma T, Vette A, Whitaker D, and Williams RW. An  
705 evaluation of EPA s National-Scale Air Toxics Assessment (NATA): Comparison  
706 with benzene measurements in Detroit, Michigan. *Atmospheric Environment*.  
707 2011;45:3301-8.
- 708 63. Xue Z and Jia C. A Model-to-Monitor Evaluation of 2011 National-Scale Air Toxics  
709 Assessment (NATA). *Toxics*. 2019;7(1).
- 710 64. Garcia E, Hurley S, Nelson DO, Gunier RB, Hertz A, and Reynolds P. Evaluation of  
711 the agreement between modeled and monitored ambient hazardous air pollutants in  
712 California. *International journal of environmental health research*.  
713 2014;24(4):363-77.
- 714 65. United States Department of Health and Human Services National Institute of Health,  
715 National Heart Lung and Blood Institute. National Asthma Education and Prevention  
716 Program Expert Panel Report 3: Guidelines for the Diagnosis and Management of  
717 Asthma. NIH Publication Number 08-5846. National Heart Lung and Blood Institute  
718 (US), ed. Clinical Practice Guidelines. Bethesda (MD). 2007.
- 719 66. Bettiol A, Gelain E, Milanese E, Asta F, and Rusconi F. The first 1000 days of life:  
720 traffic-related air pollution and development of wheezing and asthma in childhood. A  
721 systematic review of birth cohort studies. *Environmental health : a global access  
722 science source*. 2021;20(1):46.
- 723 67. Gehring U, Wijga AH, Koppelman GH, Vonk JM, Smit HA, and Brunekreef B. Air  
724 pollution and the development of asthma from birth until young adulthood. *Eur  
725 Respir J*. 2020;56(1).
- 726 68. Hsu HH, Chiu YH, Coull BA, Kloog I, Schwartz J, Lee A, et al. Prenatal Particulate  
727 Air Pollution and Asthma Onset in Urban Children. Identifying Sensitive Windows  
728 and Sex Differences. *Am J Respir Crit Care Med*. 2015;192(9):1052-9.
- 729 69. Weitekamp Chelsea A, Lein M, Strum M, Morris M, Palma T, Smith D, et al. An  
730 Examination of National Cancer Risk Based on Monitored Hazardous Air Pollutants.  
731 *Environmental health perspectives*. 2021;129(3):037008.
- 732 70. United States Census Bureau. United States Census 2016. ZIP Codes Business  
733 Patterns 2016 Data. Available at  
734 <https://www2.census.gov/programs-surveys/cbp/datasets/2016/>: Accessed 5/28/2020.
- 735 71. Oliphant TE. Python for Scientific Computing. *Computing in Science & Engineering*.  
736 2007;9(3):10-20.
- 737 72. Pedregosa F, Varoquaux G, Gramfort A, Michel V, Thirion B, Grisel O, et al.  
738 Scikit-learn: Machine Learning in Python. *Journal of Machine Learning Research*.  
739 2011;12:2825-30.
- 740 73. Seabold S, and Perktold J. Proceedings of the 9th Python in Science Conference.  
741 *Austin, TX*. 2010:61.



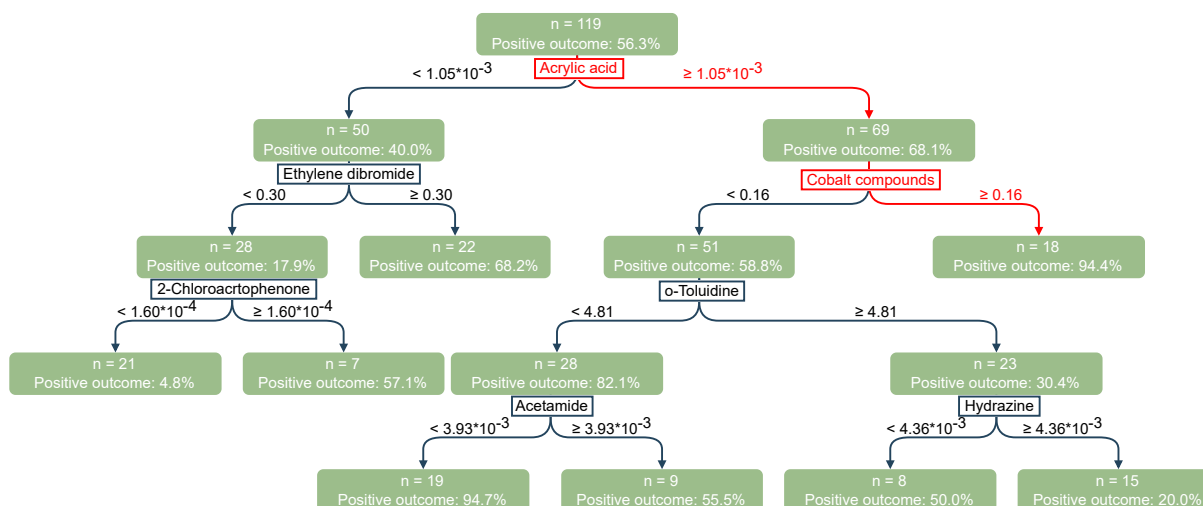
**Figure 1: Study overview.** Exposure data for over a hundred air toxics from the US Environmental Protection Agency’s National Air Toxic Assessment (NATA) database were geocoded to Airway in Asthma (ARIA) cohort participants with mild to severe persistent asthma ( $n=151$ ), based on participants’ residential ZIP code. The Data-driven Exposure Profile extraction (DEEP) method developed in this study was then applied to the air toxic data to identify multi-air toxic combinations associated with three childhood asthma outcomes: use of prescribed daily asthma controller medication, lifetime emergency department visit for asthma, and lifetime overnight hospitalization for asthma. In the first stage of DEEP, multi-air toxic combinations were identified via XGBoost models consisting of decision trees. In the second stage, multivariable logistic regression models were used to identify air toxic combinations significantly associated with childhood asthma outcomes after adjustment for age, gender, race/ethnicity, and family income. (Some images in this figure were obtained from [www.flaticon.com](http://www.flaticon.com) and were made by Wanicon, Freepik and Flat Icons.)



**Figure 2: Air toxics individually associated with childhood asthma outcomes after adjustment for age, gender, race/ethnicity, and family income in ARIA cohort participants with persistent asthma (n=151).** For each outcome and air toxic, the strength of the association is shown in terms of its odds ratio (OR), 95% confidence interval (CI), and false discovery rate (FDR). P-values for individual air toxics were obtained from multivariable logistic regression models and then adjusted for multiple hypothesis testing using the Benjamini-Hochberg procedure, yielding FDR values.

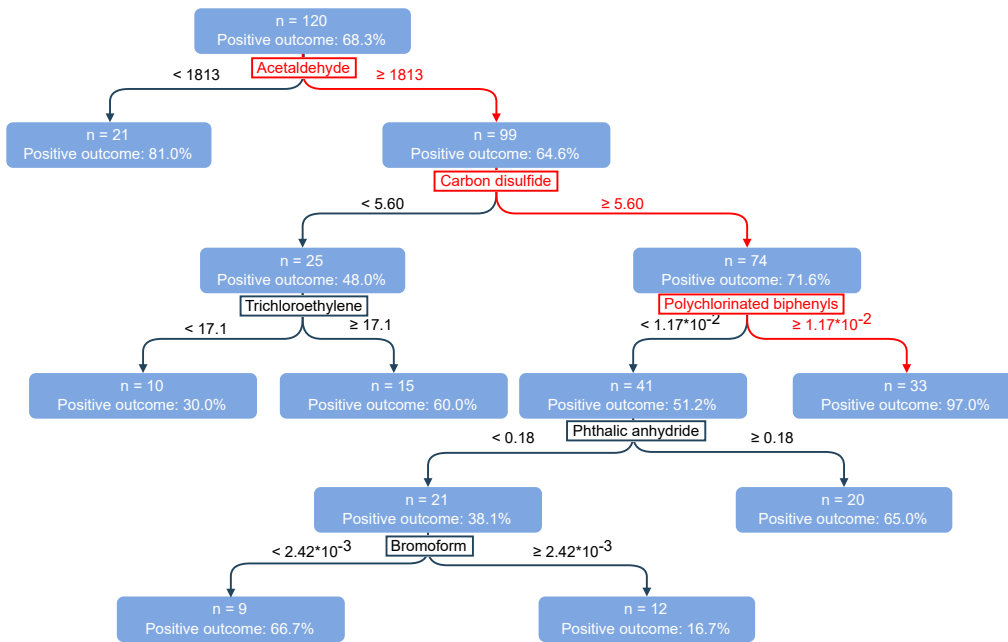


**Figure 3: Multi-air toxic combinations associated with childhood asthma outcomes after adjustment for age, gender, race/ethnicity, and family income in ARIA cohort participants with persistent asthma (n=151).** For each outcome and combination, the strength of the association is shown in terms of its odds ratio (OR), 95% confidence interval (CI), and false discovery rate (FDR). The P-values for multi-air toxic combinations were obtained from multivariable logistic regression models and then adjusted for multiple hypothesis testing using the Benjamini-Hochberg procedure, yielding FDR values.

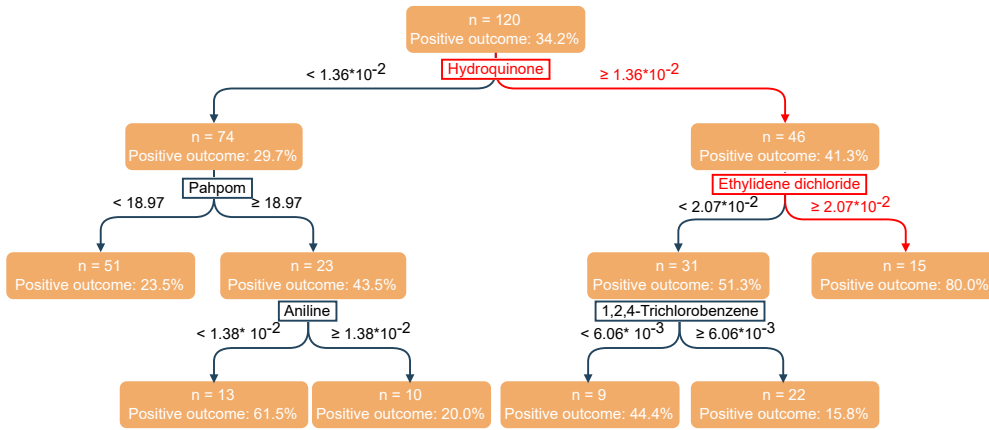


**Figure 4: A sample decision tree learned by DEEP to predict daily asthma controller medication using NATA-derived air toxic data geocoded to subjects (n=149).** Each node in the tree indicates the number of participants satisfying the air toxic decision path until that point and the percentage of participants with that outcome. The sample corresponding to each node is stratified into two subpopulations based on the air toxic and its threshold associated with the node. The multi-air toxic combination acrylic acid & cobalt compounds, which was the most significantly associated with this outcome, is highlighted in red.





**Figure 5: A sample decision tree learned by DEEP to predict lifetime emergency room visit for asthma from NATA-derived air toxic exposure data geocoded to each subject (n=151).** Each node in the tree indicates the number of participants satisfying the air toxic decision path until that point and the percentage of participants with that outcome. The sample corresponding to each node is stratified into two subpopulations based on the air toxic and its threshold associated with the node. The multi-air toxic combination acetaldehyde & carbon disulfide & polychlorinated biphenyls, which was the most significantly associated with this outcome, is highlighted in red.



**Figure 6: A sample decision tree learned by DEEP to predict lifetime overnight hospitalization for asthma from NATA-derived air toxic data geocoded to each participant (n=151).** Each node in the tree indicates the number of participants satisfying the air toxic decision path until that point and the percentage of subjects with that outcome. The sample corresponding to each node is stratified into two subpopulations based on the air toxic and its threshold associated with the node. The multi-air toxic combination hydroquinone & ethylidene dichloride, which was the most significantly associated with this outcome, is highlighted in red.

**Table 1: Characteristics of the ARIA cohort participants included in this study.** Mean (standard deviation) and number (%) are shown for continuous and categorical variables respectively. Characteristics stratified by asthma outcome are also shown. P-values comparing subjects with and without the outcome are shown from two-sided student t-test for continuous variables and a chi-squared test for categorical variables. \*Two participants did not provide information about their daily controller medication use.

	All	Daily Controller Medication*			Emergency Room Visit			Overnight Hospitalization		
	(n=151)	Yes (n=84)	No (n=65)	p-value	Yes (n=103)	No (n=48)	p-value	Yes (n=51)	No (n=100)	p-value
Age, years	12.0 (3.2)	11.5 (3.2)	12.6 (3.1)	0.048	11.7 (3.3)	12.4 (2.9)	0.23	12.1 (3.2)	11.9 (3.2)	0.69
Gender, female	62 (41.1%)	33 (39.3%)	29 (44.6%)	0.63	41 (39.8%)	21 (43.8%)	0.78	19 (37.3%)	43 (43%)	0.61
Race/ethnicity				0.50			0.03			0.07
Asian	6 (4.0%)	4 (4.8%)	2 (3.1%)		3 (2.9%)	3 (6.3%)		1 (2.0%)	5 (5.0%)	
Black	26 (17.2%)	13 (15.5%)	13 (20%)		20 (19.4%)	6 (12.5%)		12 (23.5%)	14 (14.0%)	
White	57 (37.7%)	33 (39.3%)	24 (36.9%)		32 (31.1%)	25 (52.1%)		12 (23.5%)	45 (45.0%)	
Latino	51 (33.8%)	25 (29.8%)	24 (36.9%)		42 (40.8%)	9 (18.8%)		23 (45.1%)	28 (28.0%)	
Mixed	10 (6.6%)	8 (9.5)	2 (3.1%)		5 (4.9%)	5 (10.4%)		3 (5.9%)	7 (7.0%)	
Not Reported	1 (0.7%)	1 (1.2%)	0 (0%)		1 (1.0%)	0 (0.0%)		0 (0%)	1 (1.0%)	
Income, US\$	49349.3 (18725.8)	48954.8 (18701.7)	49961.7 (19017.3)	0.75	48870.9 (19414.5)	50375.8 (17109.8)	0.63	45576.7 (14822.9)	51273.3 (20161.2)	0.053
Asthma Control Test (ACT) score	16.8 (3.9)	16.4 (3.9)	17.3 (4.0)	0.16	16.2 (4.0)	18.0 (3.5)	9.54*10 <sup>-3</sup>	15.9 (4.3)	17.3 (3.7)	0.05
FEV1%	87.1 (17.5)	85.5 (17.7)	88.6 (17.3)	0.32	86.2 (18.0)	88.6 (16.6)	0.45	82.1 (19.7)	89.4 (15.9)	0.04
FEV1/FVC	79.6 (10.4)	78.3 (10.5)	80.8 (10.2)	0.17	79.4 (10.9)	80.0 (9.5)	0.77	77.7 (10.5)	80.4 (10.3)	0.17
Regular use of asthma medicine										
Beta agonist	145 (96.0%)	83 (98.8%)	60 (92.3%)	0.11	102 (99.0%)	43 (89.6%)	0.02	51 (100%)	94 (94.0%)	0.18
Inhaled corticosteroid (ICS)	39 (25.8%)	35 (41.7%)	3 (4.6%)	7.19*10 <sup>-7</sup>	27 (26.2%)	12 (25.0%)	1	10 (19.6%)	29 (29.0%)	0.29
Combined ICS/LABA	31 (20.5%)	30 (35.7%)	0 (0%)	2.16*10 <sup>-7</sup>	28 (27.2%)	3 (6.3%)	5.97*10 <sup>-3</sup>	23 (45.1%)	8 (8.0%)	2.98*10 <sup>-7</sup>
Leukotriene receptor antagonist	30 (19.9%)	29 (34.5%)	1 (1.5%)	1.81*10 <sup>-6</sup>	23 (22.3%)	7 (14.6%)	0.37	19 (37.3%)	11 (11.0%)	3.08*10 <sup>-4</sup>
Omalizumab	2 (1.3%)	1 (1.2%)	1 (1.5%)	1	2 (1.9%)	0 (0.0%)	0.84	0 (0.0%)	2 (2.0%)	0.79

**Table 2: Air toxics identified by DEEP as significantly associated with at least one of the three asthma outcomes, either individually or in combination with other air toxics.** D: Air toxic associated with daily asthma controller medication. E: Air toxic associated with lifetime emergency room visit for asthma. O: Air toxic associated with lifetime overnight hospitalization for asthma.

Air toxic	Level at participant's residential ZIP code Median (IQR) values (in ng/m <sup>3</sup> )	Individually significant	Significant in combination(s)
1,1,1-trichloroethane	667.7 (120.5, 1452.9)		D
1,3-dichloropropene	129 (1.75, 405.18)		D
1,4-dioxane	3.16x10 <sup>-3</sup> (1.3x10 <sup>-3</sup> , 0.18)		O
2,4-toluene diisocyanate	0.08 (7.2x10 <sup>-3</sup> , 0.15)		D
Acetaldehyde	2420.6 (1967.0, 2768.3)		E
Acetamide	2.61x10 <sup>-4</sup> (1.7x10 <sup>-4</sup> , 9.1x10 <sup>-4</sup> )	D, O	
Acetophenone	0.1 (0.02, 0.17)		D
Acrylamide	4.24x10 <sup>-4</sup> (2.5x10 <sup>-6</sup> , 6.6x10 <sup>-3</sup> )		E
Acrylic acid	1.27x10 <sup>-3</sup> (4.98x10 <sup>-4</sup> , 0.06)	D	D, E, O
Arsenic compounds	0.73 (0.49, 1.13)	O	
Benzidine	1.44x10 <sup>-5</sup> (1.2x10 <sup>-5</sup> , 2.5x10 <sup>-5</sup> )	D	
Beryllium compounds	0.16 (0.09, 0.27)	E	
Carbon disulfide	5.66 (5.59, 9.57)		E
Carbonyl sulfide	0.03 (1.77x10 <sup>-3</sup> , 0.25)		O
Chlorine	0.91 (0.39, 1.28)	D	
Cobalt compounds	0.15 (0.10, 0.19)		D
Cresol cresylic acid	20.05 (15.1, 24.9)	D	
Dimethyl phthalate	0.07 (0.04, 0.08)		D
Ethyl chloride	8.21 (7.8x10 <sup>-3</sup> , 20.26)	O	D
Ethylidene dichloride	0.12 (0.02, 0.22)		O
Glycol ethers	12.67 (2.58, 117.36)		O
Hexachlorobenzene	1.2x10 <sup>-6</sup> (2.47x10 <sup>-7</sup> , 3.91x10 <sup>-4</sup> )	D	D
Hydrochloric acid	208.8 (146.8, 301.7)	O	
Hydroquinone	7.3x10 <sup>-3</sup> (1.4x10 <sup>-3</sup> , 0.03)	O	E,O
Methyl methacrylate	1.62 (1.06, 2.79)		O
Mercury compounds	0.4 (0.22, 1.87)	E	E
Pentachlorophenol	4.6x10 <sup>-7</sup> (0, 1.4x10 <sup>-6</sup> )	D, E	
Phenol	3.34 (2.03, 22.35)	D	
Phosphorus	0.21 (0.11, 0.31)		D
Polychlorinated biphenyls	1.1x10 <sup>-4</sup> (1.2x10 <sup>-5</sup> , 8.8x10 <sup>-3</sup> )	D,E	E
Quinoline	0.14 (5.23x10 <sup>-4</sup> , 0.17)	O	
Toluene	9392 (4984.58, 21249.3)		D
Trichloroethylene	106.4 (18.1, 132.6)	O	O
Triethylamine	2.81 (1.82, 5.77)	O	

**Table 3: Air toxic combinations associated with asthma outcomes with statistically significant interactions between combination members.** The p-values are for the interaction term of multivariable logistic regression models where asthma outcome was the dependent variable and independent variables included member air toxic levels, interaction term, and covariates (age, gender, race/ethnicity, family income).

<b>Outcome</b>	<b>Air Toxic Combination</b>	<b>Interactions</b>	<b>P-value</b>
Daily Controller Medication	Acrylic acid & Dimethyl phthalate	Acrylic acid & Dimethyl phthalate	0.02
	Acrylic acid & Cobalt compounds	Acrylic acid & Cobalt compounds	0.02
Emergency Room Visit	Acrylic acid & Hydroquinone	Acrylic acid & Hydroquinone	< 0.004

**Table 4: Demographic characteristics of children exposed and not exposed to the acrylic acid & cobalt compounds combination, which was associated with daily asthma controller medication.** Mean (standard deviation) and number (%) are shown for continuous and categorical variables, respectively. Also shown are p-values for the differences between the two groups of children, calculated using a two-sided two sample student t-test for continuous variables and a chi-square test for categorical variables.

Variable	Multi-air toxic combination: Acrylic acid & Cobalt compounds			p-value
	Total (n=149)	Children exposed to combination (n=23)	Children not exposed to combination (n=126)	
Age, years	12.0 (3.2)	13.7 (3.6)	11.7 (3.0)	0.02
Gender, female	87 (58.4)	14 (60.9)	73 (57.9)	0.97
Race/ethnicity				0.089
Asian	6 (4.0)	0 (0.0)	6 (4.8)	
Black	26 (17.4)	5 (21.7)	21 (16.7)	
White	57 (38.3)	6 (26.1)	51 (40.5)	
Latino	49 (32.9)	8 (34.8)	41 (32.5)	
Mixed	10 (6.7)	3 (13.0)	7 (5.6)	
Not Reported	1 (0.7)	1 (4.3)	0 (0.0)	
Income, US\$	49394.0 (18846.6)	46339.2 (15671.1)	49951.7 (19318.1)	0.34

**Table 5: Demographic characteristics of children exposed and not exposed to the acetaldehyde & carbon disulfide & polychlorinated biphenyls combination, which was associated with lifetime emergency room visit for asthma.** Mean (standard deviation) and number (%) are shown for continuous and categorical variables, respectively. Also shown are p-values for the differences between the two groups of children, calculated using a two-sided student t-test for continuous variables and a chi-square test for categorical variables.

Variable	Multi-air toxic combination: Acetaldehyde & Carbon disulfide & Polychlorinated biphenyls			p-value
	Total (n=151)	Children exposed to combination (n=48)	Children not exposed to combination (n=103)	
Age, years	12.0 (3.2)	10.0 (2.6)	12.9 (3.0)	5.34x10 <sup>-8</sup>
Gender, female	89 (58.9)	28 (58.3)	61 (59.2)	0.94
Race/ethnicity				0.0046
Asian	6 (4.0)	3 (6.2)	3 (2.9)	
Black	26 (17.2)	14 (29.2)	12 (11.7)	
White	57 (37.7)	9 (18.8)	48 (46.6)	
Latino	51 (33.8)	19 (39.6)	32 (31.1)	
Mixed	10 (6.6)	2 (4.2)	8 (7.8)	
Not Reported	1 (0.7)	1 (2.1)	0 (0.0)	
Income, US\$	49349.3 (18725.8)	44708.3 (14015.1)	51512.0 (20194.5)	0.019

**Table 6: Demographic characteristics of children exposed and not exposed to the hydroquinone & ethylidene dichloride combination, which was associated with lifetime overnight hospitalization for asthma.** Mean (standard deviation) and number (%) are shown for the continuous and categorical variables, respectively. Also shown are p-values for the differences between the two groups of children, calculated using a two-sided student t-test for continuous variables and a chi-squared test for categorical variables.

<b>Multi-air toxic combination: Hydroquinone &amp; Ethylidene dichloride</b>				
<b>Variable</b>	<b>Total (n=151)</b>	<b>Children exposed to combination (n=17)</b>	<b>Children not exposed to combination (n=134)</b>	<b>p-value</b>
Age, years	12.0 (3.2)	9.9 (2.5)	12.2 (3.2)	0.0022
Gender, female	89 (58.9)	12 (70.6)	77 (57.5)	0.44
Race/ethnicity				0.32
Asian	6 (4.0)	0 (0.0)	6 (4.5)	
Black	26 (17.2)	2 (11.8)	24 (17.9)	
White	57 (37.7)	4 (23.5)	53 (39.6)	
Latino	51 (33.8)	10 (58.8)	41 (30.6)	
Mixed	10 (6.6)	1 (5.9)	9 (6.7)	
Not Reported	1 (0.7)	0 (0.0)	1 (0.7)	
Income, US\$	49349.3 (18725.8)	40226.8 (6929.4)	50506.6 (19420.4)	8.26x10 <sup>-5</sup>